2012 Benthic Macroinvertebrate Monitoring El Dorado Hydroelectric Project (FERC 184)

Prepared for:

2890 Mosquito Road Placerville, CA 95667

Prepared by:

Garcia and Associates 1 Saunders Ave San Anselmo, CA 94960

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INTRODUCTION

Pursuant to the El Dorado Hydroelectric Project (Project 184) Benthic Macroinvertebrate Monitoring Plan (Plan; GANDA 2010), El Dorado Irrigation District (District) is required conduct BMI monitoring in various Project-affected and reference stream reaches throughout Project 184 watersheds. Per the Plan, bioassessment surveys are required during the first two years of each five-year period of the current Project 184 License (including 2011 and 2012). BMI monitoring efforts conducted during the Project 184 relicensing process between 1999 and 2001 (ECORP 2002) helped establish the Project's ecological resource objective for BMIs which states that macroinvertebrate indices (metrics) in Project-affected reaches should be similar to those in reference reaches located within and outside of the South Fork American River (SFAR) and Upper Truckee River (UTR) drainages.

Previous bioassessment surveys conducted in the Project 184 area followed the California Stream Bioassessment Procedure (CSBP) originally developed by the California Department of Fish Game (CDFG 2003). The Project 184 license requires macroinvertebrate monitoring using the CSBP method or such method as revised in the future. The current accepted methodology is the State's Surface Water Ambient Monitoring Program (SWAMP) Standard Operating Procedures for Collecting Benthic Macroinvertebrate Samples and Associated Physical and Chemical Data for Ambient Bioassessments in California (SWAMP 2007) which officially replaced the CSBP as the statewide standard for ambient bioassessment in 2007. Therefore, the SWAMP bioassessment protocol is the methodology specified by the Plan.

The District tasked Garcia and Associates (GANDA) to conduct 2012 BMI bioassessment surveys in Project 184 watersheds. This report presents the results of SWAMP bioassessment surveys conducted as specified in the Plan during fall 2012.

METHODS

Site Selection

The Plan specifies monitoring at a total of 18 sites in Project-affected reaches and associated reference reaches within Project 184 watersheds. These watersheds include the following (some of which contain paired sites located above and below existing diversion points):

- Echo Creek (Site EC-B1)
- Pyramid Creek (Site PY-B1)
- Caples Creek (Site CA-B1)
- Silver Fork American River (Site SV-B2)
- South Fork American River (Site SO-B1)
- No Name Creek (Sites NN-B1 and NN-B2)
- Alder Creek (Sites AR-B1 and AR-B2)
- Bull Creek (Sites BU-B1 and BU-B2)
- Ogilby Creek (Sites OG-B1 and OB-B2)
- Esmeralda Creek (Sites ES-B1 and ES-B2)
- Strawberry Creek (Site SB-B1)
- Sherman Canyon Creek (Site SH-B1)
- Woods Creek (Site WC-B1)

The 18 bioassessment sites are located in the same Project-affected and reference reaches specified in the Plan (see Figure 1). GPS locations for each site are listed in Table 1. Generally, 2011 and 2012 SWAMP bioassessment sites were located as close as possible to those sites selected previously during 1999-2001 relicensing efforts (ECORP 2002), although specific site boundaries for SWAMP survey reaches were established by GANDA field crews in 2011 that may be slightly upstream or downstream from the original areas sampled under the CSBP (for example, because the SWAMP protocol requires a longer survey reach than the CSBP). All sites sampled in 2012 were identical to those sampled in 2011.

Benthic Macroinvertebrate Sampling

Teams of two to four GANDA biologists conducted all benthic macroinvertebrate sampling following the SWAMP protocol. Field sampling was performed between October 8 and November 6, 2012. Sites consisted of 150-meter survey reaches wherever possible. Consistent with SWAMP protocol, shorter survey reaches were established at smaller tributaries including Esmeralda Creek (ES-B1 and ES-B2), No Name Creek (NN-B1 and NN-B2) and Ogilby Creek (OG-B1 and OG-B2) in order to avoid barriers or other confounding areas (*e.g.,* steep waterfalls, cliff areas, culverts, *etc.*). At each of these smaller tributary sites, there were numerous pool-riffle sequences to sample within the established survey reach. For larger streams (wetted width greater than 20 m), SWAMP protocol recommends increasing site length. There was one site where wetted width was consistently greater than 20 meters (Site SO-B1 on the South Fork American River [SFAR] below Kyburz Diversion Dam). However, the total survey reach length was not increased at this site because sufficient representative habitat was present within the 150-m reach and extending the site would have only added large, deep pool habitat that could not be sampled.

At sites located at elevations below 6,500 feet (PY-B1, SO-B1, NN-B1 and 2, AR-B1 and 2, BU-B1 and 2, OG-B1 and 2, ES-B1 and 2, SB-B1, SH-B1), BMI samples were collected as reachwide benthos (RWB) samples. RWB samples were compilations of eleven 1-ft² kick samples collected at the 11 main transects comprising the SWAMP survey reach. At sites near or above 6,500 feet (EC-B1, CA-B1, SV-B2, WC-B1), BMI samples were collected as both RWB samples and targeted riffle composite (TRC) samples. RWB samples were collected as described above; TRC samples were compilations of eight 1-ft² kick samples collected at eight randomly selected riffle locations within each SWAMP survey reach. Decisions regarding which sample types to collect at which locations were made by the District in consultation with CDFG's SWAMP bioassessment coordinator.

All benthic samples were collected using a Wildco® 18-by-9- inch stream-bottom sampler fitted with a 0.5 mm (500 micron) mesh bag. Samples were collected from downstream to upstream before physical habitat measurements to prevent excessive bottom trampling. At sites where both types of samples were collected, TRC and RWB samples were collected simultaneously in two separate nets while moving from downstream to upstream between transects. All samples were elutriated and cleaned in the field, placed in jars, labeled, and preserved in 10 percent formalin.

Physical Habitat Characterization

Physical habitat parameters (bankfull and wetted width, bankfull height, water depth, substrate composition, cobble embeddedness, algal cover, riparian vegetation, instream habitat complexity, canopy cover, human influence, bank stability, *etc*.) were evaluated at a combination of 11 primary and 10 secondary cross-sectional transects located along the survey reach. The "full" level of effort for physical habitat characterization as described in the SWAMP protocol was performed at all sites. Stream gradient at each site was measured using a clinometer and stadia rod (with eye-level marked) positioned at water's surface from transect to transect; compass bearings between transect mid-points were also measured. The upper, middle and lower portions of each SWAMP survey reach were documented with photographs taken in both the upstream and downstream directions, and both ends of each survey reach were marked using GPS.

Discharge was measured using the standard USGS 20-point velocity-area method at all sites where stream gage data was not available; for streams where depths and velocities were too shallow and slow to measure flows in this manner, discharge was estimated using the buoyant object method to estimate surface velocities.

Basic *in situ* water quality measurements were also taken at each site. Measured parameters included water temperature, pH, specific conductance, and dissolved oxygen concentration. All water quality measurements were collected prior to benthic macroinvertebrate sampling efforts at each site.

Laboratory Protocol

All benthic samples were processed and identified by Jon Lee Consulting. The laboratory subsampling procedure allowed separation of large/rare specimens from finer subsampled material so that more accurate estimations of the whole-sample taxa lists could be made. All samples were subsampled to a minimum of 600 individuals, although the last grid section $(i.e.,$ the aliquot containing the $600th$ individual) was always picked through and identified in its entirety to allow accurate estimation of the total sample abundance (and thus benthic density); therefore, in practice typically 625-675 organisms were identified in the laboratory. This higher level of effort (identifying a minimum of 600 instead of 500 individuals from each sample) is recommended to insure that closer to 500 clearly identifiable specimens are achieved after excluding any ambiguous and/or immature specimens. All specimens were identified to Level II standard taxonomic effort (STE) as defined by the Southwest Association of Freshwater Invertebrate Taxonomists (SAFIT), which generally corresponds to

the genus-species level for most insects, and slightly less rigorous effort (*e.g.,* class, family, or tribe/subfamily) for certain other taxa groups (Level II STE for California taxa is defined in SAFIT [2006]).

Data Analysis

Summary metrics for each replicate sample were calculated using a Microsoft Access database. Metrics are measurable attributes of macroinvertebrate communities that are known to change in response to disturbance or impairment of the stream environment. Metrics included standard richness, composition, tolerance/intolerance, and functional feeding group measures (see Table 2). All sample metrics were calculated from 500-organism fixed-count samples generated from the complete laboratory-identified taxa lists for each sample (500-count taxa lists are the standard for calculating metrics). Sample data were randomly re-sampled and standardized in this manner to achieve uniformity in count between all samples for comparative analyses (*e.g.,* so that the total number of taxa would be accurately represented for each site at a standardized level of effort, regardless of how many organisms were originally identified in the laboratory from each different sample).

In order to reduce the complexity of the information contained in the numerous metrics that describe each sample, data were compiled into a single multi-metric index, the Hydropower Index of Biotic Integrity, or Hydropower-IBI (Rehn 2010). This IBI was developed by the California Department of Fish and Game (CDFG) Aquatic Bioassessment Laboratory to be sensitive to the cumulative effects of hydropower operations on stream benthic communities. The seven component metrics of the Hydropower-IBI (ET taxa richness, %intolerant individuals, %scrapers, %non-insect taxa, Shannon diversity, %predators, and %tolerant individuals) were chosen from over 80 candidate metrics calculated using a combined dataset from nine separate studies of regulated rivers in California managed for hydropower. Values for these constituent metrics were scored (0-10) according to specific thresholds (defined in Table 3) and final Hydropower-IBI scores were achieved by summing the constituent metric scores and adjusting the index to a 100-point scale. Note that although this IBI was originally developed using only TRC-type samples, IBI scores were calculated for both TRC and RWB samples for all Project 184 SWAMP data because recent published and unpublished analyses suggest that RWB and TRC methods can produce generally comparable results across a broad range of settings within California (Van Buuren and Ode 2008). Therefore, it was assumed that RWB samples collected during this study contained sufficient riffle material for Hydropower-IBI analysis. Further details regarding development of the Hydropower-IBI are provided in Rehn (2010).

Ten percent of the benthic macroinvertebrate samples collected in 2011 and 2012 (2 randomly selected samples out of the 22 total samples collected each year) were submitted to CDFG's Aquatic Bioassessment Laboratory for an independent quality assurance/quality control (QA/QC) check for accuracy of enumeration and taxonomic identification.

RESULTS

2012 Benthic Macroinvertebrate Summary

In 2012, it is estimated that nearly 145,000 benthic macroinvertebrates were collected from the 18 sites in the Project 184 area (in TRC and RWB samples combined). Of these individuals, 15,113 specimens were identified, representing 221 different taxa from 72 families and 15 taxonomic orders (per SAFIT Level II STE). The most common taxa included clinger mayflies of the genus *Cinygmula,* the nemourid stonefly *Zapada cinctipes,* mayflies of the genus *Paraleptophlebia,* stoneflies of the genus *Sweltsa*, caddisflies of the genus *Lepidostoma*, elmid beetles of the genus *Heterlimnius,* and aquatic earthworms of the class Oligochaeta. Other common taxa included brachycentrid mayflies of the genus *Micrasema*, chironomid midges of the genus *Micropsectra*, blackflies of the genus *Simulium*, and mayflies of the genera *Ephemerella* and *Ironodes.* Complete taxa lists for 500-organism fixed-counts and estimated whole-sample taxa lists for all samples are presented in Appendices A and B, respectively.

The average number of taxa per sample for all sites (including both TRC and RWB samples) was 47, including an average of 23 EPT taxa. Shannon Diversity averaged 2.95 and Shannon Evenness average 0.77. Percent EPT averaged 64 percent (46% of which were sensitive EPT) and the dominant taxon comprised 12 percent of the average sample. Tolerant and intolerant individuals comprised 2 and 44 percent of the average sample, respectively. The mean weighted tolerance value was 3.2. On average, collectors were the dominant functional feeding group (33%), followed by shredders (19%), scrapers (18%), predators (17%), filterers (7%), macrophyte herbivores (4%), omnivores (1%), and piercer herbivores (<1%). Macroinvertebrate density averaged 596 individuals/ft² for all samples. A summary of biological metrics for 500-organism fixed-counts from all TRC and RWB samples is presented in Table 4. Results of the CDFG laboratory's taxonomic QA/QC check will be reported (as they become available) if any significant discrepancies are found.

2012 Physical Habitat/Water Quality Summary

SWAMP bioassessment sites surveyed in the Project 184 area in 2012 ranked between "optimal" and "marginal" in terms of available epifaunal substrate and cover, sediment deposition, and channel alteration (rapid bioassessment [RPB] scores are ranked by category as poor, marginal, suboptimal, or optimal). Stream gradient ranged from low (1.6% slope at Caples Creek) to very high (28.0% slope at upper Bull Creek). Human influences encountered in the vicinity of survey reaches included rip-rapped banks, cabins, roads, diversion pipes, campgrounds, historical logging, and bridge abutments (defined as "walls/rip-rap/dams," "buildings," "roads/railroads," "pipes (inlet/outlet)," "park/lawn," "logging operations," and 'bridges/abutments" on the SWAMP survey form, respectively).

Water quality parameters measured at SWAMP bioassessment sites were within acceptable ranges during the fall 2012 SWAMP surveys, with water temperatures ranging from 4.8 to 13.4°C, pH ranging from 6.9 to 7.9, and dissolved oxygen concentration ranging from 9.4 to 15.6 milligrams/liter. Discharge ranged from less than 1.0 cubic foot per second (cfs) in several smaller creeks to 33.5 cfs in the mainstem SFAR during our surveys. A summary of physical habitat data and water quality measurements collected at each site in 2012 is presented in Tables 5 and 6 (Table 5 summarizes reach-wide habitat measurements collected once at each SWAMP site and Tables 6a through 6c summarize transect-based measurements collected at multiple cross-sections within each SWAMP survey reach). Site photographs are compiled in Appendix C. Copies of original SWAMP field datasheets are provided in Appendix D.

DISCUSSION

Comparisons between Reference Reaches and Project-Affected Reaches (2012)

Overall, samples collected from Project-affected reaches scored slightly lower on average in terms of certain richness, composition, tolerance, and functional feeding group measures than those collected from reference reaches during 2012 SWAMP surveys (Table 4). Although some variation was apparent among individual metrics and samples, scores for the multimetric Hydropower-IBI averaged 16 percent higher overall in references reaches (52) than Project-affected reaches (45) (Figure 2).

Total taxa richness averaged 11 percent higher in reference reaches versus Project-affected reaches (50 vs. 45 total taxa, respectively). Richness of individual samples ranged from 61 taxa collected in the RWB sample from upper Bull Creek (Site BU-B2), to 27 taxa collected in the RWB sample from Caples Creek below Caples Lake (Site CA-B1). Shannon Diversity averaged seven percent higher at reference sites versus Project-affected sites (3.06 vs. 2.85, respectively; see Figure 3). Diversity of individual samples ranged from 3.56 in the RWB sample from upper upper Bull Creek (Site BU-B2), to 1.93 in the TRC sample from Woods Creek (Site WC-B1; note that this site was intermittent during summer 2012 prior to fall sampling). Macroinvertebrate density was slightly lower on average in reference reaches than Project-affected reaches (552 vs. 633 individuals/ft², respectively). Among individual samples, density was lowest in the RWB sample from Woods Creek above Caples Lake (Site WC-B1) (78 individuals/ft²) and highest in the TRC sample from Caples Creek below Caples Lake (Site CA-B1) (2,061 individuals/ft²).

Composition measures were more variable overall among reference and Project-affected sites. Average values for most composition measures were similar for reference and Projectaffected reaches (Table 4). The overall percentage of insects was high for all samples (82% to 100%).

The average percent composition of tolerant organisms was very low for all samples (0% to 6%) and the average percent composition of intolerant organisms was typically high (21% to 80%). Thus, average weighted tolerance values were relatively low (*i.e.,* good) for both reference and Project-affected reaches.

Functional feeding group measures were similar overall among reference and Projectaffected reaches, however samples from reference reaches had more scrapers (28% vs. 12%) and fewer filterers (4% vs. 10%) on average (Table 4).

The average composition of the major taxonomic groups differed slightly among reference reaches and Project-affected reaches in 2012. In terms of the major insect orders, mayflies (Order Ephemeroptera), stoneflies (Order Plecoptera), and beetles (Order Coleoptera) were more abundant on average in samples from reference reaches, whereas caddisflies (Order Trichoptera) and true flies (Order Diptera) were more abundant in samples from Projectaffected reaches (Figure 4). Non-insect taxa were much less abundant overall than insects, although snails (Class Gastropoda) were more abundant on average in samples from reference reaches, while aquatic earthworms (Class Oligochaeta), freshwater mites (Class Acari), and clams (Order Bivalvia) were more abundant in samples from Project-affected reaches (Figure 5).

Comparisons between 2011 and 2012 Survey Results

Results of 2011 and 2012 surveys were very similar overall. Individual metrics (richness, composition, tolerance/intolerance, and functional feeding group measures) as well as multimetric hydropower IBI scores from each year were similar both on a site-by-site and a Project-vs.-reference reach basis. IBI scores were slightly lower in 2012 than in 2011 (45 and 52 in 2012 vs. 49 and 57 in 2011 for Project-affected and reference reaches, respectively- see Figure 6). It appears that these small differences in IBI scores, along with other slight between-year differences in individual metrics, are well within the range of potential natural (inter-annual) variability. As such, 2011 and 2012 bioassessment data adequately characterize the existing biological and physical habitat conditions in Project watersheds for the current five-year operational period, while providing valuable baseline information for comparisons with future bioassessment data.

CONCLUSIONS

Bioassessment data collected in 2011 and 2012 indicate that Project 184 watersheds generally support relatively robust benthic communities (in terms of richness, composition, tolerance, and functional feeding group measures) characterized by good overall water quality. Physical habitat conditions were predominantly in the optimal to suboptimal range. Overall, these data suggest that no major problems with biological integrity are occurring in Project watersheds.

The ecological resource objective for benthic macroinvertebrates (as defined in the Appendix B, Section 1 of the El Dorado Relicensing Settlement Agreement) states that "macroinvertebrate indices (metrics) in Project-affected stream reaches should be similar to reference reaches." In 2011 and 2012, Project-affected reaches scored slightly lower on average than reference reaches in terms of many individual metrics as well as the multimetric Hydropower-IBI. While some variability was evident between sites, most richness

measures (total taxa richness, Shannon diversity and evenness, *etc*.) were higher on average at reference sites than Project-affected sites in 2011 and 2012. However, some composition measures (*e.g.,* %sensitive EPT), tolerance measures (*e.g.,* %intolerant individuals), and functional feeding group measures (*e.g.,* %shredders) averaged higher at Project-affected sites than at reference sites in these years. To a certain extent, IBI scores and component metric values would be expected to be higher at unregulated vs. regulated sites (*e.g.,* %scrapers was consistently higher at reference sites than at Project-affected sites in 2011 and 2012); and indeed, overall Hydropower-IBI scores were 17 and 16 percent higher for reference reaches than Project-affected reaches in 2011 and 2012, respectively. Furthermore, most reference sites are located nearer to headwater reaches where biological integrity tends to be naturally higher than in downstream reaches where most Project-affected sites are located.

As such, it is likely that many of the observed differences in metric averages between Projectaffected reaches and reference reaches primarily reflect ecological differences between upstream and downstream locations (*i.e.,* underlying differences in stream hydrology, substrate, morphology, gradient, riparian influences, *etc.*) rather than Project-related differences. Given the existing variability in summary metrics and the potential measurement of mostly ecological differences, it is reasonable to conclude that the narrative ecological resources objective defined in Appendix B, Section 1 of the El Dorado Relicensing Settlement Agreement is currently being met (*i.e.,* Project-affected reaches and reference reaches are adequately "similar" overall).

RECOMMENDATIONS

Over the course of conducting the 2011 and 2012 bioassessment work, GANDA has developed the following recommendations for consideration in future monitoring efforts:

- Numerical ecological resource objectives should be developed based on 2011 and 2012 data to replace the current narrative objective.
- Better reference sites are needed for certain paired Project-affected sites (*i.e.,* Ogilby Creek, Caples/Woods Creek, Alder Creek) such that bioassessments may better isolate Project-related differences as opposed to simply measuring underlying ecological differences. Currently, such paired comparisons are not ecologically valid due to inherent differences in stream hydrology, substrate, morphology, gradient, and riparian influences between upstream and downstream sites. For example, upper Ogilby Creek (Site OG-B2) is consistently dry with zero surface flow for most of each summer, whereas lower Ogilby Creek (Site OG-B1) is perennial; Woods Creek (Site WC-B1) is a steep, headwater stream that becomes intermittent in low snowpack years, whereas Caples Creek is a low-gradient and higher-order perennial stream; upper Alder Creek (Site AR-B2 near the headwaters) is nearly three miles upstream of lower Alder Creek (Site AR-B1) which has much different stream morphology, gradient, substrate, and site elevation.
- Although 2011 and 2012 sampling was scheduled to correspond to the timing of previous relicensing surveys (October-November), future sampling should be conducted earlier in the season (*e.g.,* August-September as opposed to October-November) so as to conform to the standard index sampling period recommended by SWAMP (as well as to avoid access issues associated with the potential for early snowstorms later in the fall). Index periods help standardize sampling during the most stable flow periods in order to minimize variation in the biological communities being sampled (the index period for the Project 184 area is June-September). Sampling in September versus October, for example, would not appreciably affect summary metrics since benthic communities are relatively stable during these lowflow periods and metric scoring tools are generally very robust to such small seasonal differences. Therefore, September sampling would represent a reasonable compromise between conforming to the standard index period and maintaining maximum comparability with data previously collected primarily during the month of October.
- The collection of TRC samples should be omitted as RWB samples alone will suffice. SWAMP continues to focus on RWB samples only and the initial modification of the protocol to target riffles in order to ensure adequate representation of the benthos at steeper, higher elevation sites for this Project does not appear necessary.
- When sampling in consecutive years, physical habitat data collection could be minimized in the second year (although benthic samples and water quality data should continue to be collected at each site both years). Most of these sites are characterized by very stable stream morphology such that channel aggradation/degradation or meander is unlikely. Thus, perhaps only a subset of transects (*e.g.,* transects A, F, and K only) could be re-measured the second year to verify key aspects of the physical habitat characterization as opposed to repeating the full effort (labor associated with full physical habitat measurements comprises the vast majority of all field labor during a standard SWAMP effort). If conditions appear to have changed from one year to the next at a given site (*e.g.,* due to a landslide, bank failure, or other erosive event), the full level of effort could be repeated at that site to capture such local changes.

REFERENCES

- California Department of Fish and Game (CDFG). 2003*.* California Stream Bioassessment Procedure (Protocol Brief for Biological and Physical/Habitat Assessment in Wadeable Streams). CDFG Aquatic Bioassessment Laboratory, revised December 2003. 11 pp.
- ECORP Consulting (ECORP). 2002. Draft Benthic Macroinvertebrate Sampling Program: El Dorado Irrigation District, Hydroelectric Project 184 (El Dorado County, California). Dated 4 June 2002.
- Garcia and Associates (GANDA). 2010. El Dorado Hydroelectric Project FERC Project No. 184 Benthic Macroinvertebrate Monitoring Plan. Version 2.0. Prepared by GANDA, December 2010.
- Rehn, A.C. 2010. Benthic Macroinvertebrates as Indicators of Biological Condition below Hydropower Dams. California Energy Commission, PIER Energy‐Related Environmental Research Program. CEC‐500‐2009‐060.
- Southwest Association of Freshwater Invertebrate Taxonomists (SAFIT). 2006. List of Freshwater Macroinvertebrate Taxa from California and Adjacent States including Standard Taxonomic Effort Levels. Second circulating draft, dated 27 September 2006.
- Surface Water Ambient Monitoring Program (SWAMP). 2007. Standard Operating Procedures for Collecting Benthic Macroinvertebrate Samples and Associated Physical and Chemical Data for Ambient Bioassessments in California. February 2007.
- Van Buuren, B. H., and P. Ode. 2008. Surface Water Ambient Monitoring Program Quality Assurance Program Memorandum. Memo to SWAMP Roundtable, dated 17 September 2008.

FIGURE 2. Multi-metric Hydropower-IBI scores in Project-affected vs. reference reaches

FIGURE 3. Benthic community diversity in Project-affected vs. reference reaches

FIGURE 4. Abundance of major insect orders in Project-affected vs. reference reaches

FIGURE 5. Abundance of major non-insect classes in Project-affected vs. reference reaches

FIGURE 6. Multi-metric Hydropower-IBI scores in Project-affected vs. reference reaches (2011-2012)

SITE ID	DESCRIPTION	UTM LOCATION ¹	
		Upstream ²	Downstream ²
AR-B1	Alder Creek below diversion	10 S 0727817 4293722	10 S 0727783 4293846
$AR-B2$	Alder Creek above diversion	10 S 0730155 4291030	10 S 0730155 4291140
BU-B1	Bull Creek below diversion	10 S 0723080 4294280	10 S 0722997 4294368
BU-B ₂	Bull Creek above diversion	10 S 0723612 4293646	10 S 0723542 4293736
$CA-B1$	Caples Creek below Caples Lake	10 S 0756345 4288557	10 S 0756231 4288551
EC-B1	Echo Creek below Lower Echo Lake	10 S 0757821 4303759	10 S 0757934 4303807
ES-B1	Esmerelda Creek below diversion	10 S 0718115 4293217	10 S 0718078 4293288
ES-B ₂	Esmerelda Creek above diversion	10 S 0718332 4292992	10 S 0718311 4293066
$NN-B1$	No Name Creek below diversion	10 S 0731140 4293874	10 S 0731124 4293956
$NN-B2$	No Name Creek above diversion	10 S 0731173 4293746	10 S 0731153 4293794
OG-B1	Ogilby Creek below diversion	10 S 0718893 4293859	10 S 0718909 4293906
OG-B2	Ogilby Creek above diversion	10 S 0720413 4293075	10 S 0720346 4293141
PY-B1	Pyramid Creek below Lake Aloha	10 S 0750292 4300308	10 S 0750294 4300162
$SB-B1$	Strawberry Creek near SFAR confluence	10 S 0747420 4296859	10 S 0747312 4296920
$SH-B1$	Sherman Canyon Creek	10 S 0743689 4285807	10 S 0743619 4285914
$SO-B1$	South Fork American below Kyburz diversion	10 S 0732883 4294117	10 S 0732748 4294072
$SV-B2$	Silver Fork American below Silver Lake	10 S 0750229 4284442	10 S 0750132 4284527
$WC-B1$	Woods Creek above Caples Lake	10 S 0758190 4287291	10 S 0758071 4287309

TABLE 1. GPS locations of 2012 SWAMP bioassessment survey reaches in Project 184 area.

¹ GPS datum: NAD 83; ² Upsream and downstream locations are endpoints of each SWAMP survey reach (corresponding to main survey transects "K" and "A," *respectively).*

TABLE 3. Scoring ranges for constituent metrics of the Hydropower-IBI. Thresholds shown are for 500-organism fixed-count samples identified to SAFIT Level II standard taxonomic effort (after Rehn 2010).

TABLE 4. Summary of biological metrics for 2012 Project 184 SWAMP bioassessment samples

**Dominant land use= forest (F), subutrb/town (S), rangeland (R)*

TABLE 6a. Summary of transect-based physical habitat measurements from 2011 Project 184 SWAMP bioassessment sites

TABLE 6b. Summary of Transect-Based Physical Habitat Measurements from 2011 Project 184 SWAMP Bioassessment Sites (cont'd)

**Habitat Complexity Codes= Absent (A), Sparse (S), Moderate (M), Heavy (H), Very Heavy (VH)*

TABLE 6c. Summary of Transect-Based Physical Habitat Measurements from 2011 Project 184 SWAMP Bioassessment Sites (cont'd)

**Riparian Vegetation Codes= Absent (A), Sparse (S), Moderate (M), Heavy (H), Very Heavy (VH)*

Appendix A

500-Organism Fixed-Count Taxa Lists 2012 Project 184 SWAMP Bioassessment

Appendix B

Estimated Whole-Sample Taxa Lists 2012 Project 184 SWAMP Bioassessment

Appendix C

Site Photographs 2012 Project 184 SWAMP Bioassessment

FIGURE AR-B1-1. Looking upstream from the bottom transect (A) at Site AR-B1

FIGURE AR-B1-2. Looking downstream from the bottom transect (A) at Site AR-B1

FIGURE AR-B1-3. Looking upstream from the middle transect (F) at Site AR-B1

FIGURE AR-B1-4. Looking downstream from the middle transect (F) at Site AR-B1

FIGURE AR-B1-5. Looking upstream from the upper transect (K) at Site AR-B1

FIGURE AR-B1-6. Looking downstream from the upper transect (K) at Site AR-B1

FIGURE AR-B2-1. Looking upstream from the bottom transect (A) at Site AR-B2

FIGURE AR-B2-2. Looking downstream from the bottom transect (A) at Site AR-B2

FIGURE AR-B2-3. Looking upstream from the middle transect (F) at Site AR-B2

FIGURE AR-B2-4. Looking downstream from the middle transect (F) at Site AR-B2

FIGURE AR-B2-5. Looking upstream from the upper transect (K) at Site AR-B2

FIGURE AR-B2-6. Looking downstream from the upper transect (K) at Site AR-B2

FIGURE BU-B1-1. Looking upstream from the bottom transect (A) at Site BU-B1

FIGURE BU-B1-2. Looking downstream from the bottom transect (A) at Site BU-B1

FIGURE BU-B1-3. Looking upstream from the middle transect (F) at Site BU-B1

FIGURE BU-B1-4. Looking downstream from the middle transect (F) at Site BU-B1

FIGURE BU-B1-5. Looking upstream from the upper transect (K) at Site BU-B1

FIGURE BU-B1-6. Looking downstream from the upper transect (K) at Site BU-B1

FIGURE BU-B2-1. Looking upstream from the bottom transect (A) at Site BU-B2

FIGURE BU-B2-2. Looking downstream from the bottom transect (A) at Site BU-B2

FIGURE BU-B2-3. Looking upstream from the middle transect (F) at Site BU-B2

FIGURE BU-B2-4. Looking downstream from the middle transect (F) at Site BU-B2

FIGURE BU-B2-5. Looking upstream from the upper transect (K) at Site BU-B2

FIGURE BU-B2-6. Looking downstream from the upper transect (K) at Site BU-B2

FIGURE CA-B1-1. Looking upstream from the bottom transect (A) at Site CA-B1

FIGURE CA-B1-2. Looking downstream from the bottom transect (A) at Site CA-B1

FIGURE CA-B1-3. Looking upstream from the middle transect (F) at Site CA-B1

FIGURE CA-B1-4. Looking downstream from the middle transect (F) at Site CA-B1

FIGURE CA-B1-5. Looking upstream from the upper transect (K) at Site CA-B1

FIGURE CA-B1-6. Looking downstream from the upper transect (K) at Site CA-B1

FIGURE EC-B1-1. Looking upstream from the bottom transect (A) at Site EC-B1

FIGURE EC-B1-2. Looking downstream from the bottom transect (A) at Site EC-B1

FIGURE EC-B1-3. Looking upstream from the middle transect (F) at Site EC-B1

FIGURE EC-B1-4. Looking downstream from the middle transect (F) at Site EC-B1

FIGURE EC-B1-5. Looking upstream from the upper transect (K) at Site EC-B1

FIGURE EC-B1-6. Looking downstream from the upper transect (K) at Site EC-B1

FIGURE ES-B1-1. Looking upstream from the bottom transect (A) at Site ES-B1

FIGURE ES-B1-2. Looking downstream from the bottom transect (A) at Site ES-B1

FIGURE ES-B1-3. Looking upstream from the middle transect (F) at Site ES-B1

FIGURE ES-B1-4. Looking downstream from the middle transect (F) at Site ES-B1

FIGURE ES-B1-5. Looking upstream from the upper transect (K) at Site ES-B1

FIGURE ES-B1-6. Looking downstream from the upper transect (K) at Site ES-B1

FIGURE ES-B2-1. Looking upstream from the bottom transect (A) at Site ES-B2

FIGURE ES-B2-2. Looking downstream from the bottom transect (A) at Site ES-B2

FIGURE ES-B2-3. Looking upstream from the middle transect (F) at Site ES-B2

FIGURE ES-B2-4. Looking downstream from the middle transect (F) at Site ES-B2

FIGURE ES-B2-5. Looking upstream from the upper transect (K) at Site ES-B2

FIGURE ES-B2-6. Looking downstream from the upper transect (K) at Site ES-B2

FIGURE NN-B1-1. Looking upstream from the bottom transect (A) at Site NN-B1

FIGURE NN-B1-2. Looking downstream from the bottom transect (A) at Site NN-B1

FIGURE NN-B1-3. Looking upstream from the middle transect (F) at Site NN-B1

FIGURE NN-B1-4. Looking downstream from the middle transect (F) at Site NN-B1

FIGURE NN-B1-5. Looking upstream from the upper transect (K) at Site NN-B1

FIGURE NN-B1-6. Looking downstream from the upper transect (K) at Site NN-B1

FIGURE NN-B2-1. Looking upstream from the bottom transect (A) at Site NN-B2

FIGURE NN-B2-2. Looking downstream from the bottom transect (A) at Site NN-B2

FIGURE NN-B2-3. Looking upstream from the middle transect (F) at Site NN-B2

FIGURE NN-B2-4. Looking downstream from the middle transect (F) at Site NN-B2

FIGURE NN-B2-5. Looking upstream from the upper transect (K) at Site NN-B2

FIGURE NN-B2-6. Looking downstream from the upper transect (K) at Site NN-B2

FIGURE OG-B1-1. Looking upstream from the bottom transect (A) at Site OG-B1

FIGURE OG-B1-2. Looking downstream from the bottom transect (A) at Site OG-B1

FIGURE OG-B1-3. Looking upstream from the middle transect (F) at Site OG-B1

FIGURE OG-B1-4. Looking downstream from the middle transect (F) at Site OG-B1

FIGURE OG-B1-5. Looking upstream from the upper transect (K) at Site OG-B1

FIGURE OG-B1-6. Looking downstream from the upper transect (K) at Site OG-B1

FIGURE OG-B2-1. Looking upstream from the bottom transect (A) at Site OG-B2

FIGURE OG-B2-2. Looking downstream from the bottom transect (A) at Site OG-B2

FIGURE OG-B2-3. Looking upstream from the middle transect (F) at Site OG-B2

FIGURE OG-B2-4. Looking downstream from the middle transect (F) at Site OG-B2

FIGURE OG-B2-5. Looking upstream from the upper transect (K) at Site OG-B2

FIGURE OG-B2-6. Looking downstream from the upper transect (K) at Site OG-B2

FIGURE PY-B1-1. Looking upstream from the bottom transect (A) at Site PY-B1

FIGURE PY-B1-2. Looking downstream from the bottom transect (A) at Site PY-B1

FIGURE PY-B1-3. Looking upstream from the middle transect (F) at Site PY-B1

FIGURE PY-B1-4. Looking downstream from the middle transect (F) at Site PY-B1

FIGURE PY-B1-5. Looking upstream from the upper transect (K) at Site PY-B1

FIGURE PY-B1-6. Looking downstream from the upper transect (K) at Site PY-B1

FIGURE SB-B1-1. Looking upstream from the bottom transect (A) at Site SB-B1

FIGURE SB-B1-2. Looking downstream from the bottom transect (A) at Site SB-B1

FIGURE SB-B1-3. Looking upstream from the middle transect (F) at Site SB-B1

FIGURE SB-B1-4. Looking downstream from the middle transect (F) at Site SB-B1

FIGURE SB-B1-5. Looking upstream from the upper transect (K) at Site SB-B1

FIGURE SB-B1-6. Looking downstream from the upper transect (K) at Site SB-B1

FIGURE SH-B1-1. Looking upstream from the bottom transect (A) at Site SH-B1

FIGURE SH-B1-2. Looking downstream from the bottom transect (A) at Site SH-B1

FIGURE SH-B1-3. Looking upstream from the middle transect (F) at Site SH-B1

FIGURE SH-B1-4. Looking downstream from the middle transect (F) at Site SH-B1

FIGURE SH-B1-5. Looking upstream from the upper transect (K) at Site SH-B1

FIGURE SH-B1-6. Looking downstream from the upper transect (K) at Site SH-B1

FIGURE SO-B1-1. Looking upstream from the bottom transect (A) at Site SO-B1

FIGURE SO-B1-2. Looking downstream from the bottom transect (A) at Site SO-B1

FIGURE SO-B1-3. Looking upstream from the middle transect (F) at Site SO-B1

FIGURE SO-B1-4. Looking downstream from the middle transect (F) at Site SO-B1

FIGURE SO-B1-5. Looking upstream from the upper transect (K) at Site SO-B1

FIGURE SO-B1-6. Looking downstream from the upper transect (K) at Site SO-B1

FIGURE SV-B2-1. Looking upstream from the bottom transect (A) at Site SV-B2

FIGURE SV-B2-2. Looking downstream from the bottom transect (A) at Site SV-B2

FIGURE SV-B2-3. Looking upstream from the middle transect (F) at Site SV-B2

FIGURE SV-B2-4. Looking downstream from the middle transect (F) at Site SV-B2

FIGURE SV-B2-5. Looking upstream from the upper transect (K) at Site SV-B2

FIGURE SV-B2-6. Looking downstream from the upper transect (K) at Site SV-B2

FIGURE WC-B1-1. Looking upstream from the bottom transect (A) at Site WC-B1

FIGURE WC-B1-2. Looking downstream from the bottom transect (A) at Site WC-B1

FIGURE WC-B1-3. Looking upstream from the middle transect (F) at Site WC-B1

FIGURE WC-B1-4. Looking downstream from the middle transect (F) at Site WC-B1

FIGURE WC-B1-5. Looking upstream from the upper transect (K) at Site WC-B1

FIGURE WC-B1-6. Looking downstream from the upper transect (K) at Site WC-B1

Appendix D

Copies of Field Datasheets 2012 Project 184 SWAMP Bioassessment